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8 Fine-scale population structure of rainbow trout (*Oncorhynchus mykiss*) in the  
9 Spokane River drainage in relation to hatchery stocking and barriers  
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27 Abstract: We examined population structure in rainbow trout (*Oncorhynchus mykiss*)  
28 collected from 20 tributaries and three mainstems in the greater Spokane River drainage  
29 using 13 microsatellite loci. Populations displayed some excess homozygosity and  
30 linkage disequilibrium, which was more pronounced in upper tributary collections and  
31 likely the result of small effective population sizes or structuring within tributaries. In  
32 general, population structure followed geographic structure: collections from creeks  
33 within sub-drainages were most closely related and collections from different tributaries  
34 were genetically distinct. Comparisons to cutthroat trout (*O. clarki*) indicated little to no  
35 introgression. Comparisons to steelhead, coastal, and inland rainbow trout from  
36 hatcheries suggested introgression by hatchery fish into some wild populations.  
37 Introgression was suspected in populations from stocked tributaries and where the  
38 tributary lacked barriers to escaped hatchery fish. Populations from tributaries above  
39 barriers that had not been stocked were genetically distinct from hatchery fish and  
40 appeared to be native inland redband rainbow trout.

41

41 Introduction: Effective fisheries management is based upon an understanding of  
42 population structure, usually a complex reflection of historical processes, geography and  
43 life history (Small *et al.* 1998; McCusker *et al.* 2000; Spidle *et al.* 2003; Spruell *et al.*  
44 2003). Historically, tributaries in the Pacific Northwest were recolonized following  
45 glacial retreat, with population structure reflecting common founders (Small *et al.* 1998,  
46 McCusker *et al.* 2000). Since salmonids home to their natal stream for breeding, genetic  
47 structure is often organized upon geographic structure of drainages (McCusker *et al.*  
48 2000, Taylor *et al.* 2003). As some amount of straying naturally occurs within drainages  
49 and to a lesser extent among drainages within the same region, population structure  
50 generally follows a hierarchy of regional structure with populations more closely related  
51 in nearby drainages (Hansen and Mensberg 1998; Spruell *et al.* 2003; Taylor *et al.* 2003).  
52 In addition to natural movement, fisheries managers move trout among drainages and  
53 among regions, potentially altering genetic structure. Hatchery introductions have mixed  
54 impacts upon wild populations: salmonids are regionally adapted (Taylor 1991) and  
55 hatchery fish, often of non-local origin, may lack characteristics allowing them to  
56 succeed in regions different from their origins or to succeed under natural conditions  
57 (Currens *et al.* 1990; Hindar *et al.* 1991; Williams *et al.* 1996; Reisenbichler and Rubin  
58 1999; Hansen *et al.* 2001; Weber and Fausch 2003; McGinnity *et al.* 2004).  
59 Anthropogenic barriers to fish movement also impact population structure (Neraas and  
60 Spruell 2001; Van Houdt *et al.* 2005). Dams, culverts or periodic loss of flow within a  
61 waterway from land-use practices that prevent fish from moving throughout drainages  
62 may affect gene flow and biodiversity patterns (Taylor *et al.* 2003), and lead to smaller  
63 effective population sizes. Natural barriers such as historic dry channels or waterfalls

64 arising from tectonic processes similarly influence fish movement (Currens *et al.* 1990),  
65 but might leave a different genetic signature where barriers predated humans. Salmonid  
66 management is based partly upon understanding the relative roles of natural and  
67 anthropogenic influences on salmonid ecology as reflected in population genetic  
68 structure.

69

70 In this study, population genetic structure was investigated in rainbow trout occupying  
71 tributaries of the Spokane and Little Spokane rivers in eastern Washington State (Figure  
72 1) using microsatellite DNA. We tested the hypothesis that hatchery fish planted in the  
73 system over the past century had displaced or introgressed into native populations and  
74 examined the influence of natural and anthropogenic barriers on putative native  
75 population structure. Native inland “redband” rainbow trout (*Oncorhynchus mykiss*  
76 *gairdneri*), both anadromous (steelhead) and resident forms, were once abundant. Dam  
77 construction in the early 1900s on the Spokane River (Figure 1) eliminated the  
78 anadromous form but native resident populations may still persist in unstocked tributaries  
79 isolated by barriers. To mitigate for dam construction, hatchery rainbow trout and  
80 hatchery cutthroat trout (*O. clarki*) were stocked throughout the drainage starting in the  
81 early 1900s. Hatchery rainbow trout were primarily coastal rainbow trout (*O. mykiss*  
82 *irideus*) but also included some inland redband rainbow trout, (inland rainbow trout are  
83 distinguished morphologically from coastal rainbow trout by a red lateral stripe and  
84 primitive taxonomic structures (Currens *et al.* 1990), genetically by differences at  
85 allozyme loci (Currens 1997) and ecologically by adaptations to inland environments  
86 (Currens 1997)). Stocking efforts varied by tributary and were based upon availability

87 of lakes, habitat quality and potential for recreational opportunities. We found that  
88 hatchery fish had introgressed into native populations in tributaries with hatchery  
89 stocking and in unstocked tributaries below barriers that were exposed to escaped  
90 hatchery fish. Unstocked tributaries above barriers appeared to be native inland redband  
91 rainbow trout. Thus, although barriers disrupted connectivity among rainbow trout  
92 populations, barriers also served to protect native diversity.

93

#### 94 Materials and methods:

##### 95 Area geography

96 The Spokane River system is a low gradient drainage underlain by a large aquifer  
97 maintained by snow-melt and rainwater. Although only upper headwaters were within  
98 Pleistocene glacial margins, glacial Lake Spokane covered part of the drainage and part  
99 of the area was scoured during the Great Spokane flood around 18,000 years before  
100 present (USGS 2006). In addition to waterfalls from tectonic and glacial activity, and  
101 man-made barriers (culverts, dams, high sediment load), movement of fish among  
102 tributaries is restricted by availability of water since in many creeks some portion is dry  
103 for at least part of the year. Although springs throughout the drainage provide high  
104 quality water and thermal refugia, water availability and quality in some tributaries has  
105 declined dramatically over the past century as riparian plants were removed and road  
106 building, mining, timber harvest and agriculture diverted groundwater and increased  
107 water temperature, sedimentation and pollution (Bruce Kinkead, tribal biologist, Coeur  
108 d'Alene Tribe, pers. comm.).

109

110 Samples and area history

111 Samples of adult fish fin tissue were obtained non-lethally by backpack electrofishing in  
112 a stratified random sampling design (10 fish per 100 m blocked section). Rainbow trout  
113 samples were collected 2001 through 2004 from 21 tributaries and the mainstem of the  
114 Spokane (356 fish) and Little Spokane (940 fish) rivers and Hangman Creek (206  
115 rainbow and 108 cutthroat; Figure 1, Table 1). In most tributaries with barriers, samples  
116 were collected above barriers (Table 1). Some barriers were complete and one-way (eg.  
117 Spokane Falls), others were incomplete (eg. culvert surmountable under high flow) and  
118 others were absent under rare conditions (eg. 5.5 km of dry creek bed in each of Coulee  
119 and Deep creeks flows only in years of high snowfall). In Dartford Cr., samples were  
120 gathered above and below a culvert about 200 m from the creek mouth.

121

122 Historically, inland resident (rainbow) and anadromous (steelhead) rainbow trout  
123 were throughout the system and cutthroat trout were only above Spokane Falls (Behnke  
124 1992). Hatchery rainbow trout have been planted since the early 1900's. Most fish  
125 originated at the Spokane Hatchery (Figure 1, Table 2), which maintains a coastal  
126 rainbow trout broodstock from McCloud River, CA (Crawford 1979; Busack and Gall  
127 1980; Nielsen *et al.* 1999; 96 fish sample included). Other hatchery rainbow trout  
128 planted included Trout Lodge Inc. Hatchery (private hatchery, broodstock origin  
129 unknown, 49 fish sample), and Phalon Lake Conservation stock (Phalon Lake), an inland  
130 redband rainbow stock from Kettle River 100 km north of Spokane River. The Phalon  
131 Lake stock was reconstructed in 2002 and we included samples from 2001 and 2002 (200  
132 fish). Some Trout Lodge fish were planted by landowners and some escaped from ponds

133 in upper Deep Cr. (Jason McLellan, WDFW, unpublished data). Three coastal rainbow  
134 trout hatchery stocks, Goldendale, Eells Springs and South Tacoma (McCloud River  
135 origin with additional components), while not planted into the system or part of the study,  
136 were included in the cluster analyses to strengthen the examination of relationships  
137 between wild-origin inland rainbow trout and hatchery-origin coastal rainbow trout.  
138 Some Lyons Ferry Hatchery steelhead (native anadromous inland rainbow trout,  
139 broodstock origin mainly in upper Columbia River; Bumgarner *et al.* 2003), were also  
140 planted (100 fish included). Introgression by hatchery rainbow trout might be detected as  
141 clustering with the hatchery sample rather than with populations from nearby tributaries  
142 and by coastal ancestry signals in inland populations.

143

144 Hatchery cutthroat trout, planted less extensively (Table 2), were King's Lake broodstock  
145 from the Pend Oreille drainage. Nehchen Cr. received at least unofficial one cutthroat  
146 infusion around 20 years ago when a landowner brought in fish following dewatering of  
147 the creek (Bruce Kinhead, Coeur d'Alene Tribal Biologist, pers. comm.). Two Nehchen  
148 Cr. samples (102 fish) were included to determine if fish were cutthroat trout rather than  
149 rainbow trout (redband rainbow have markings similar to cutthroat, Behnke 1992). We  
150 compared all samples in the study to three cutthroat trout samples from the Pend Oreille  
151 drainage that had been collected for another study, (Sullivan Lake, Sullivan Cr., and Gold  
152 Cr.), to identify potential cutthroat trout and possible hybrids. Prior to stocking (and  
153 introduction) cutthroat trout were only present above Spokane Falls. Cutthroat and  
154 rainbow trout may hybridize, but naturally occupy different portions of a tributary  
155 (Allendorf *et al.* 2001; Marshall *et al.* 2006). Fish may be forced into proximity and thus

156 hybridize if hatchery cutthroat trout are introduced into a rainbow-occupied tributary with  
157 limited habitat.

158

#### 159 Sample processing

160 DNA from 1502 wild-origin rainbow trout, 445 hatchery rainbow trout and 102 wild-  
161 origin cutthroat trout was extracted using a chelex protocol (Small *et al.* 1998).

162 Microsatellite alleles at 14 loci were PCR-amplified using fluorescently labeled primers

163 (see Table 3 for detailed PCR information). PCR's were conducted on an MJResearch

164 PTC-200 thermocycler in 96 well plates in 5  $\mu$ l volumes employing 1  $\mu$ l template with

165 final concentrations of 1.5 mM MgCl<sub>2</sub>, 200 $\mu$ M of each dNTP, 1X Promega PCR buffer,

166 and 0.01U Taq polymerase. After initial three minute denature at 92°, 33 cycles

167 consisting of 92° for 15 seconds, annealing (temperature in Table 3) for 30 seconds,

168 extension at 72° for 60 seconds were followed by a 30 minute extension at 72°. Samples

169 were run on ABI 3100 and 3730 automated sequencers and alleles were sized (to base

170 pairs) and binned using an internal lane size standard (GS500Liz from Applied

171 Biosystems) and Genemapper software (Applied Biosystems). A subsample was run on

172 both sequencer platforms to standardize allele mobilities.

173

#### 174 Within collection data analysis

175 Statistical tests were conducted on loci and samples from each collection site to assess

176 conformation to Hardy Weinberg expectations (Hardy Weinberg equilibrium, HWE), and

177 genotypic heterogeneity using GENEPOP version 3.3 (Raymond and Rousset 1995) with

178 100 batches and 2,000 iterations. Loci were examined for large allele drop-out, null



179 alleles and scoring errors using MICROCHECKER (Van Oosterhout *et al.* 2004). We  
180 used FSTAT 2.9.3 (Goudet 2001) to calculate  $F_{IS}$  values and their significance for each  
181 locus in each collection (350,000 randomizations), and calculate Weir and Cockerham's  
182 (1984) estimators of Wright's  $F$  statistics ( $F$ ,  $\Theta$ , and  $f$ , hereafter referred to as  $F_{IS}$ ,  $F_{IT}$ ,  
183 and  $F_{ST}$ ) and their jackknife intervals over all loci. Linkage disequilibrium was assessed  
184 using GENEPOP with 200 batches and 3,000 iterations. Allelic richness was calculated  
185 using rarefaction as implemented in HP-RARE v1.4 (Kalinowski 2004), and based upon  
186 permutation of 50 alleles. Populations were tested for population bottleneck signals  
187 using BOTTLENECK (Piry *et al.* 1998). Gene diversity (Nei 1987) was calculated using  
188 FSTAT.

189

190 Between collection data analysis

191 Partitioning of molecular variance was explored using AMOVA tests (analysis of  
192 molecular variance, Excoffier *et al.* 1992) implemented in ARLEQUIN 2.000 (Schneider  
193 *et al.* 2001) with collections organized by tributary. We conducted an assignment test  
194 using GENECLASS2 (Cornuet *et al.* 1999) to examine the likelihood that, based upon  
195 the genotype of the fish and allele frequencies in collections, an individual fish originated  
196 in the population or tributary where it was sampled. We used the Rannala and Mountain  
197 (1997) option, with collections grouped by tributary. A Bayesian analysis implemented  
198 in STRUCTURE 2.1 (Pritchard *et al.* 2000) was used to assess hybridization between  
199 rainbow and cutthroat trout, to examine introgression by hatchery rainbow trout into  
200 native populations, and to estimate individual ancestry. In initial analyses, all collections  
201 (cutthroat, hatchery and wild-origin rainbow trout) were included with  $K$  (number of

202 clusters or possible populations) set from 2 to 30. With  $K = 2$ , we hypothesized that the  
203 data set would divide into cutthroat and rainbow trout, with possible hybrid individuals  
204 sharing ancestry in both groups. With  $K = 3$ , we hypothesized that the rainbow trout  
205 cluster would divide into coastal and inland redband rainbow trout, also with hybrid  
206 individuals sharing ancestry in both groups. With  $K > 3$  we expected population  
207 structure to emerge. Initial analyses with  $K = 2$  to 5 were conducted in 10 independent  
208 runs allowing admixture with 50,000 burn-ins and 1,000,000 iterations. Due to  
209 computational intensity, only single runs (50,000 burn-ins and 1,000,000 iterations) were  
210 conducted for  $K = 6$  to 30. After identifying collections with possible hatchery  
211 introgression ( $> 25\%$  average ancestry in the cluster occupied by Spokane Hatchery  
212 collection), we conducted pairwise tests with introgressed collections and the Spokane  
213 Hatchery collection with  $K = 2$ , to look more closely at introgression (10 runs, 50,000  
214 burn-ins, 1,000,000 iterations). Other pairwise tests were employed to examine  
215 relationships among collections: collections were tested for heterogeneity in genotypic  
216 distributions at each locus and across all loci using GENEPOP with 300 batches and  
217 3,000 iterations. To examine the magnitude of differentiation between populations and  
218 between tributary groups, pairwise  $F_{ST}$  values and their significance were evaluated using  
219 FSTAT with 100,000 permutations. All test results were adjusted for multiple  
220 comparisons using sequential Bonferroni corrections.

221

222 Genetic relationships among collections were also explored with cluster analyses.  
223 Population allele frequencies were generated from genotypic data using CONVERT 1.3  
224 (Glaubitz 2004). Pairwise chord distances (Cavalli-Sforza and Edwards, 1967) among

225 collections were calculated from allele frequencies using GENDIST in PHYLIP 3.5c  
226 (Felsenstein 1993). We used geometric distance since fluctuation in population sizes  
227 should increase the impact of drift relative to mutation. Chord distances were plotted in a  
228 neighbor-joining (NJ) tree using PHYLIP. To test the repeatability of NJ tree branching,  
229 the allele frequency file was bootstrapped 10,000 times across loci using SEQBOOT.  
230 Tree topologies were created for replicates with NEIGHBOR, and a consensus tree was  
231 generated using CONSENSE and plotted with TREEVIEW 1.6.6 (Page 2001). Chord  
232 distances were also plotted in a multidimensional scaling analysis using NTSYSpC  
233 version 2.02j (Rohlf 1993) to view relationships among collections in the absence of  
234 dendrogram architecture.

235

#### 236 Results:

237

238 PCR amplifications at some loci were less successful for cutthroat trout than rainbow  
239 trout (average missing loci per individual genotype = 4 vs 2, respectively), particularly at  
240 *One-102*, *One-108*, *Ots-100*, *Ots-1*, *Omm-1130*, *Omy-1001*, and *Omy-1011*. Cutthroat  
241 trout generally had significant homozygosity at these loci (Table 4). In individual locus  
242 tests, 39 out of 462 (14 loci x 33 collections, excluding cutthroat) total  $F_{IS}$  tests were  
243 significant for excess homozygosity (Table 4, adjusted alpha,  $0.05/462 = 0.0001$ ).

244 Twelve of the significant excesses occurred at *One-108* and MICROCHECKER  
245 suggested a null allele, so this locus was excluded from further analyses.

246 MICROCHECKER indicated a possible null allele at *Omm-1130*, but since there were  
247 only five significant homozygote excesses at this locus and population HWE significance

248 values were unchanged when *Omm-1130* was removed (data not shown), we left this  
249 locus in the analysis. In global tests, one third of collections (after removing *One-108*,  
250 Table 1) were out of HWE for homozygote excess, and  $F_{IT}$  values (Table 3) indicated  
251 significant deficits of heterozygotes at all loci. Average  $F_{IS}$  values were higher in above  
252 barrier populations than below barrier populations (0.061 and 0.038, respectively), but  
253 were not significantly different ( $P = 0.06$ , FSTAT group comparison, 1000  
254 permutations).  $F_{ST}$  values indicated significant variance among populations at all loci  
255 (Table 3).

256

257 Positive  $F_{IS}$  values and disequilibrium suggested that collections contained offspring  
258 from matings among close relatives (inbreeding) within populations or that collections  
259 contained admixtures of trout from different breeding groups. To study this further, wild-  
260 origin rainbow trout collections were examined for family groups and relationships with  
261 IDENTIX (Belkhir *et al.* 2002). In the Deadman Cr. collection, the first nine individuals  
262 collected appeared to be siblings (all Q values > 0.35): Queller and Goodnight's (1989)  
263 relatedness value  $Q = 0.25$  for half-sibs and 0.5 for full-sibs. In the California Cr.  
264 collection, five individuals appeared to be full sibs. Mean relatedness values for  
265 Nehchen, Coulee, and Dartford creeks (above barrier) were in the top 5% of a random  
266 distribution of 500 permuted values. However, all but the Buck Cr. sample had  
267 significantly high variance in pairwise relatedness values among individuals, suggesting  
268 related groups within most samples (Belkhir *et al.* 2002). Further, some pairwise  
269 relatedness values suggested that half-sib relationships (offspring of mating among  
270 relatives) increased in collections from further up drainages (e.g. Lower Spokane = 2%

271 half sibs, Middle Spokane = 3.8% half sibs and Upper Spokane = 5.2% half sibs).  
272 Relatedness values were higher in above barrier collections than in below barrier  
273 collections (-0.079 and -0.129, respectively), and the differences were on the edge of  
274 significance ( $P = 0.05$ , FSTAT group test, 1000 permutations). In collections from  
275 headwaters and regions isolated above barriers, population sizes may have been restricted  
276 by available habitat, fostering mating among relatives and decreasing heterozygosity  
277 through inbreeding (Castric *et al.* 2002). In streams with intermittent good habitat,  
278 collections may have included breeding groups associated with habitat patches, which  
279 would decrease overall heterozygosity in a Wahlund-type effect (Castric *et al.* 2002).  
280 Thus, although sampling followed a randomized design, some samples contained non-  
281 random components or related individuals.

282

283 In other examinations, loci were tested for linkage in pairwise genotypic disequilibrium  
284 tests. Several collections had one to three pairs of loci (out of 78 possible pairs per  
285 population after *One-108* was removed) in linkage disequilibrium, with a high of 9 in  
286 California Cr. (Table 1). After removing four of five siblings in the California Cr.  
287 collection, only one locus pair was linked. No locus pair was linked in more than three  
288 populations. *Omm-1130* was involved in 11/29 linkages (8/21 after California Cr.  
289 siblings removed).

290

291 Among wild collections, allelic richness was highest in lower Little Spokane River and  
292 lowest in Coulee Cr. (Table 1). Allelic diversity was lower but not significantly different  
293 in collections from upper portions and smaller branches of tributaries (9.2 in above

294 barrier collections vs 10.2 in below barrier collections,  $P = 0.056$  from FSTAT group  
295 comparison, 1000 permutations) suggesting smaller effective population sizes or fewer  
296 founders in above barrier collections. Bottleneck tests conducted under the infinite allele  
297 model also supported recent reduction in population sizes in several collections (Table 1).  
298 Gene diversity generally concurred with other diversity measures: collections with higher  
299 allelic richness had higher gene diversity. Six alleles at various loci were absent from  
300 coastal hatchery collections (Spokane and Trout Lodge hatcheries) and present at  
301 intermediate frequency (4-9%) in suspected introgressed collections and present at high  
302 frequency (15 - 28%) in suspected pure redband rainbow trout collections. Sixteen  
303 alleles at various loci were found at high frequencies in coastal hatchery collections,  
304 intermediate frequencies in suspected introgressed collections, and low frequencies in  
305 suspected pure redband rainbow trout collections. Coastal hatcheries had no private  
306 alleles, Phalon Lake had 16 private alleles, and Lyons Ferry Hatchery had 10 private  
307 alleles. All alleles in coastal hatchery collections were found in wild-origin collections,  
308 although sometimes at very low frequencies (eg. 26% in Spokane Hatchery vs 0.4% in  
309 wild). Two hundred and sixteen alleles present in redband rainbow trout collections  
310 (wild or hatchery) were absent in coastal hatchery collections.

311

312 Pairwise (not shown) and global  $F_{ST}$  values (overall  $F_{ST} = 0.080$ , excluding cutthroat  
313 trout and hatchery collections) were all significantly different from 0 (pairwise test  
314 adjusted alpha =  $0.05/300 = 0.000167$ ) except within the following tributaries: Deer and  
315 Little Deer creeks, upper and lower Deep Creek, Dartford Cr. above and below barrier,  
316 2002 and 2003 middle Spokane River, and 2003 Indian and 2004 Indian from upper

317 Hangman (other Hangman collections were too small for statistical tests). Pairwise  
318 genotypic test results (not shown) were similar to pairwise  $F_{ST}$  test with one additional  
319 non-significant comparison, Little Deep and South Fork Little Deep creeks. Populations  
320 occupying tributaries within the greater Spokane drainage were thus genetically  
321 differentiated from each other and distinct from hatchery rainbow trout and steelhead.

322

323 AMOVA results supported pooling samples by tributary. Although significant variation  
324 was partitioned within tributaries (3.21%,  $P < 0.001$ ), more variation was found among  
325 tributaries (5.51%,  $P < 0.001$ ). In pairwise  $F_{ST}$  tests with collections grouped by  
326 tributary, all values were significantly different from 0 (Table 5). Coulee Cr., unstocked  
327 and isolated above a dry creek bed, was highly differentiated from other collections (see  
328 average and individual values in Table 5). Deep Cr., also isolated by a dry creek bed but  
329 exposed to hatchery escapees from ponds in upper Deep Cr. and limited stocking, was  
330 genetically closer to other collections impacted by hatchery fish through stocking or  
331 escapees: Buck and Dartford creeks and lower Little Spokane River.

332

333 The topology of the consensus neighbor-joining tree and bootstrap support illustrated  
334 associations among collections within tributaries (Figure 2). Tributary collections  
335 formed branches with at least 98% bootstrap support. Collections from lower portions of  
336 tributaries (eg. middle Spokane River) had shorter branch lengths indicating less  
337 distinction, possibly due to one-way gene flow from upper tributaries and straying from  
338 other tributaries. Cutthroat trout and Nehchen Cr. collections formed a branch with  
339 100% bootstrap support, indicating that Nehchen Cr. fish were cutthroat trout. Coastal

340 hatchery collections occupied a branch, which included Trout Lodge Hatchery, Marshall  
341 Cr., and Deep Cr. collections with 64% bootstrap support (Figure 2), indicating a coastal  
342 origin for Trout Lodge broodstock and possible hatchery introgression in Marshall and  
343 Deep creek populations. In a consensus tree with combined collections (not shown), the  
344 Buck and Dartford creek collections also joined the hatchery branch with 75% bootstrap  
345 support, also suggesting hatchery introgression into these collections. In a  
346 multidimensional scaling plot employing combined collections (Figure 3), the first axis  
347 was defined by Trout Lodge and Spokane Hatchery on the far left and Lyons Ferry  
348 Hatchery on the far right, with the Marshall, Dartford, Buck and Deep creek collections  
349 on the coastal hatchery side and putative native inland redband rainbow trout collections  
350 and Phalon Lake on the Lyons Ferry Hatchery side. The first axis appears to show a  
351 genetic gradient from coastal to introgressed inland rainbow to inland rainbow and  
352 steelhead collections (Figure 3).

353

354 Hybridization with cutthroat trout and introgression by hatchery fish were estimated with  
355 a Bayesian analysis. When the number of hypothetical populations ( $K$ ) was set at 2,  
356 cutthroat trout and Nehchen Cr. occupied one cluster and coastal and inland rainbow trout  
357 collections occupied the other cluster, with an average 3% or less cutthroat ancestry in  
358 each rainbow trout collection (Table 6). Three individual fish (individual data not  
359 shown) from upper Spokane appeared to be cutthroat-rainbow hybrids (10 - 30% ancestry  
360 in cutthroat trout cluster) and one was a possible cutthroat trout (90% ancestry in  
361 cutthroat trout cluster). With  $K = 3$  (Table 6), rainbow trout collections divided: inland  
362 redband rainbow trout collections occupied a cluster that included Lyons Ferry Hatchery



363 and Phalon Lake, and coastal hatchery groups occupied a cluster that included substantial  
364 portions (> 25%) of some inland redband rainbow trout collections that had been stocked  
365 with hatchery fish over several years (lower Little Spokane River, Buck, Deep (additional  
366 pond escapes), Dragoon, and Marshall creeks). Collections from above barriers in  
367 tributaries with single or few stockings (Otter, Deer, Coulee creeks, and Upper Little  
368 Spokane River) shared modest ancestry (6 – 18%) with coastal hatchery fish. Collections  
369 from below barriers that were stocked once or possibly exposed to hatchery strays  
370 (California, Little Deep, West Branch Dragoon and Upper Dragoon creeks) also shared  
371 modest ancestry (13 – 18%) with the coastal hatchery cluster. The mouth of Dartford Cr.  
372 was a kilometer below the Spokane Hatchery outfall, and collections from below and  
373 above the culvert (semi-permeable barrier) shared around 50% ancestry with coastal  
374 hatchery fish. Collections from unstocked tributaries above barriers (all Hangman  
375 tributaries, South Fork Deadman, North and South Fork Little Deep, and Little Deer  
376 creeks) had less than 3% ancestry in the coastal hatchery cluster. The 2001 and 2002  
377 Phalon Lake collections shared 30% and 2% ancestry with coastal hatchery fish,  
378 respectively: Phalon Lake broodstock was suspected of contamination with coastal  
379 hatchery fish and was reconstructed in 2002. In analyses with  $K = 4$  (all collections  
380 included), Spokane Hatchery and coastal hatchery collections moved into their own  
381 cluster (data not shown). With increased  $K$ , tributary groups and then single collections  
382 eventually occupied predominantly single clusters, and some collections subdivided  
383 among multiple clusters (data not shown).

384

385 We conducted further STRUCTURE analyses with suspected introgressed collections (>  
386 25% coastal ancestry) paired with the Spokane Hatchery collection with  $K = 2$  (Table 7).  
387 In these analyses inland collections shared less than 2% Spokane Hatchery ancestry.  
388 Individual fish admixture values indicated that 1% of wild-origin inland redband rainbow  
389 individuals shared more than 10% ancestry with the Spokane Hatchery collection (7/642  
390 fish, average coastal ancestry within these 7 fish = 19%, highest = 37%).

391

392 In maximum likelihood assignment tests, assignments were classified as positive and  
393 unambiguous if the assignment was 100 times more likely than the next most likely  
394 assignment, and positive but ambiguous when the assignment was less than 100 times  
395 more likely. Most fish were assigned unambiguously to their tributary of origin (Table  
396 8). Misassigned fish (fish not assigning to collection of origin) may be strays coming  
397 down out of tributaries into the Spokane and Little Spokane rivers, or fish moving among  
398 tributaries in Deadman, Little Deep and Dragoon creeks. A few fish were ambiguously  
399 misassigned between 01Phalon Lake and wild-origin rainbow collections. Since few  
400 Phalon Lake fish were planted in the system, this may reflect shared common ancestry  
401 rather than introgression. Spokane and Trout Lodge hatcheries had 100% correct  
402 assignment and no wild-origin fish were assigned to hatchery collections, suggesting low  
403 impact on native populations.

404

404 Discussion:

405 Introgression by hatchery rainbow trout

406 This study explored the population structure of wild rainbow trout populations in the  
407 greater Spokane River drainage and assessed whether hatchery fish planted throughout  
408 the system over the past 100 years had introgressed into native populations. Since we  
409 lacked genetic data for native populations prior to hatchery supplementation, we  
410 compared gene pools of wild spawning populations to gene pools maintained in  
411 hatcheries and introduced into the system within the past 60 years. A coastal strain of  
412 rainbow trout was the most extensively planted hatchery fish. Introgression by coastal  
413 hatchery rainbow trout was indicated in collections exposed to hatchery fish through  
414 escapees and planting (lower Little Spokane River, Buck, Marshall, Deep, and Dartford  
415 creeks). We suspected hatchery introgression because these collections were on the  
416 coastal branch in the neighbor-joining tree and the coastal side of the first axis in the  
417 multidimensional scaling analysis and appeared to share coastal ancestry in the large-  
418 scale Bayesian analysis. However, in contrast to other studies where some wild-origin  
419 fish assigned to hatcheries after several years of stocking (Fritzner *et al.* 2001; Hansen *et*  
420 *al.* 2001; Spidle *et al.* 2003), assignment tests suggested that hatchery impact was not as  
421 strong as might be expected. Further, other hatchery introgression studies demonstrated  
422 admixture in wild-origin individuals (Hansen 2002, Sušnik *et al.* 2004), whereas our  
423 pairwise analyses with hatchery and introgressed wild-origin indicated that wild  
424 collections shared little ancestry with hatchery collections and few individuals appeared  
425 admixed. We hypothesized that in inland redband rainbow trout populations exposed to  
426 coastal-origin hatchery rainbow, introgression shifted these gene pools away from the

427 native inland redband rainbow trout gene pool but not to the point where the individual  
428 identity of a population was compromised.

429

430 Hatchery fish have not replaced wild fish in any collection examined. All wild-origin  
431 gene pools had alleles absent from the hatchery gene pool and wild fish remain  
432 genetically distinct from hatchery fish (Table 5). All alleles in coastal-origin hatchery  
433 fish were found in inland collections, supporting introgression into inland populations:  
434 several alleles at high frequency in coastal hatchery fish were at intermediate frequency  
435 in suspected introgressed populations and at low frequency in populations unexposed to  
436 hatchery fish. Genetic drift may have eliminated some coastal-type allelic diversity  
437 during 60 years of hatchery rearing. It is unlikely that associations between wild and  
438 hatchery fish arose from native inland redband rainbow trout entering hatchery gene  
439 pools since other coastal hatchery collections clustered with Spokane and Trout Lodge  
440 hatchery collections. We suspect that, similar to other species (Williams *et al.* 1997;  
441 Nielsen *et al.* 2001; Weber and Fausch 2003; McGinnity *et al.* 2004), hatchery-adapted  
442 rainbow trout of coastal ancestry did not replace native fish because they may not have  
443 survived as well or been as successful on wild spawning grounds as regionally adapted  
444 native inland redband rainbow trout, and that hybrid offspring may have suffered reduced  
445 survival and fitness (Utter 2001).

446

447 The mainstem of the Spokane and Little Spokane rivers were heavily planted (Table 2)  
448 but showed little impact from hatchery fish. In the middle Spokane River, below  
449 Spokane Falls, hatchery fish tended to move down the river into the reservoir associated

450 with Nine Mile Falls dam (McLellan 2005a), and thus may not have used upstream  
451 mainstem and tributary spawning grounds. Similarly, the Little Spokane River enters  
452 Long Lake dam reservoir and hatchery fish may have moved down into the reservoir.  
453 Dragoon and Deadman creeks also received a fair number of hatchery fish but showed  
454 little relationship to hatchery collections. Dragoon Cr. encompassed some of the most  
455 degraded habitat in the system, although a spring in the middle of the system provided a  
456 refuge, and in Deadman Cr., good habitat was intermittent between large stretches of  
457 poor habitat (McLellan 2005a, 2005b). In both tributaries, hatchery fish may have  
458 survived poorly or were unable to compete with native fish inhabiting habitat patches.  
459  
460 Two inland stocks from Phalon Lake and Lyons Ferry Hatchery were also planted  
461 minimally into the Spokane River system. Assignment tests suggested that a few  
462 Deadman Cr., and Little Spokane River fish looked genetically like Phalon Lake fish  
463 (Table 8). Since Phalon Lake fish were planted in the Little Spokane River (Table 2),  
464 and fish from the Little Spokane River appeared to stray into Deadman Cr. (Table 8),  
465 these misassigned fish could have originated at Phalon Lake. However, three Phalon  
466 Lake fish looked genetically like wild rainbow in lower Little Spokane River. Reciprocal  
467 misassignments could also reflect recent shared ancestry: although mouths of the  
468 Spokane and Kettle rivers are around 100 km apart, prior to 1911 no barriers prevented  
469 straying. The relationship may also reflect common ancestry if colonists from the same  
470 refuge repopulated inland drainages following the retreat of glaciers and glacial lakes  
471 (Currens 1997, McCusker *et al.* 2000, Docker and Heath 2002).  
472

473 Stocking efforts were lower for cutthroat trout and we saw no evidence of genetic impact  
474 in collections from stocked regions. Four possible hybrids were collected in the upper  
475 Spokane River, which had not been stocked, but was the only region where cutthroat  
476 trout had occurred naturally prior to stocking. However, our results indicated that  
477 cutthroat trout were successfully introduced into Nehchen Cr. in the upper Hangman Cr.  
478 drainage following dewatering 20 years ago. The dewatering may have eliminated native  
479 rainbow trout, allowing the cutthroat trout unchallenged colonization. Further, these  
480 cutthroat trout were from Benewah Cr. in the Coeur d'Alene drainage across the ridge  
481 from Nehchen Cr. Ecological conditions may have been similar such that fish were  
482 adapted to conditions prevailing in Nehchen Cr.

483

484 Impact of barriers

485 The impact of hatchery supplementation was complexed with habitat quality and the  
486 presence, permeability and duration of barriers. Habitat in the Spokane drainage was  
487 affected by anthropogenic activities throughout the 1900's. In addition to hatchery  
488 supplementation and decreases in water quality and availability, dam and road building  
489 decreased physical connectivity throughout the drainage. Most collections exhibited  
490 genetic signals suggesting that populations had experienced genetic bottlenecks that  
491 might be associated with disrupted gene flow. The foremost effect of barriers upon  
492 populations depended upon whether the barrier could be traversed. All barriers allowed  
493 fish to move downstream, generating one-way gene flow between formerly connected  
494 groups of fish.

495

496 Some barriers permitted upstream movement. In Dartford Cr., the barrier was a culvert  
497 surmountable by larger fish under certain flow conditions. Although allelic richness was  
498 lower above the barrier, the collections were undifferentiated indicating that the  
499 population remained connected and possibly effectively larger, countering drift and  
500 differentiation. The semi-permeable barrier also allowed hatchery introgression into the  
501 portion of the population above the barrier.

502

503 In Deadman Cr., barriers included an unsurmountable dam as well two possibly  
504 surmountable culverts. Although genetically closest to each other, the population in SF  
505 Deadman Cr. that was isolated above the barriers differed significantly from the  
506 Deadman Cr. population below the barriers. While their genetic relationship prior to  
507 barriers is unknown, the longer branch length in the dendrogram, decreased allelic  
508 richness, and positive bottleneck signal suggested that the isolated SF Deadman Cr.  
509 population experienced enhanced genetic drift.

510

511 Within Deep Cr., a waterfall barrier predated human activity. Although the tributary was  
512 further isolated by a dry creek bed (discussed below), hatchery fish influenced collections  
513 above and below the waterfall, likely bringing the gene pools closer to each other and to  
514 other gene pools influenced by hatchery fish (Waples 1995). In the Little Spokane River,  
515 a waterfall also divided the river into upper and lower regions. While documented  
516 hatchery stocking was mild in the upper river and intense in the lower river, hatchery fish  
517 were free to move down-river into Long Lake, a preferred habitat, possibly decreasing  
518 impact on the populations. The pairwise  $F_{ST}$  value indicated that the upper and lower

519 river collections had more variance between them than all other above and below barrier  
520 comparisons, possibly because of longer isolation, fewer founders in the upper river  
521 population, and lower hatchery impact.

522

523 Habitat quality and availability also likely influenced population sizes and rates of  
524 genetic drift. Where habitat was poor or patchy and limited by a barrier, as in Dragoon  
525 Creek, drift would be stronger in smaller populations, increasing divergence seen in  
526 longer branch lengths in the dendrogram. In collections from upper drainages positive  
527  $F_{IS}$  values, higher relatedness and lower allelic richness indicated smaller population  
528 sizes. In contrast, populations below barriers where habitat was less restricted, such as  
529 the lower Spokane River and Deadman Cr., received allelic infusions from strays,  
530 increasing their allelic richness and genetic connections to other populations (where  
531 strays originated), illustrated by short branch lengths in the dendrogram and lower  
532 average pairwise  $F_{ST}$  values.

533

534 The Coulee Cr. collection was the most divergent in the study, and although habitat is  
535 good above the barrier and supported a large population, the dry creek bed isolated the  
536 tributary long before humans altered connectivity in the system. Given the episodic  
537 nature of flow throughout the full creek length, divergence may have been enhanced by a  
538 small founder group. Deep Cr. shared the barrier (and likely founders) with Coulee Cr.  
539 However, as discussed above, exposure to hatchery fish ponded above the barrier and  
540 stocked into Deep Cr. shifted the Deep Cr. gene pool closer to other hatchery-influenced  
541 collections as well as increasing allelic richness in Deep Cr.



542

543 We found that, similar to brown trout (Van Houdt *et al.* 2005), barriers protected some  
544 populations from hatchery escapees, thus preserving some native genetic diversity. In  
545 unstocked tributaries, collections from above barriers, had little to no coastal hatchery  
546 genetic signal and collections from below barriers appeared to be introgressed. The data  
547 suggested that hatchery introgression in wild populations depended upon the magnitude  
548 of exposure to hatchery fish through stocking or hatchery escapees.

549

550

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560

561

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753

753 Table 1. Statistical information for collections: the number of locus pairs in genotypic  
 754 linkage disequilibria (link), gene diversity (Gene Div), allelic richness (Rich), positive  
 755 bottleneck signal (BNeck) and  $F_{IS}$  and associated  $P$  value (bold values significant after  
 756 corrections). Combined abbreviation (Comb) indicates groups for assignment and  
 757 pairwise  $F_{ST}$  tests. "Hatch" indicates hatchery and na = sample too small. Collections  
 758 above barriers indicated by \* and stocked collections indicated by @.

River	Collection	Abbreviation	Comb	n	link	Gene Div	Rich	BNeck	$F_{IS}$	$P$ value
Upper Hangman Cr.	2003 Mission Cr.*	MSS	Hang	9	na	na	na	na	na	na
	2004 Mission Cr.*	MSS	Hang	10	na	na	na	na	na	na
	2003 Sheep Cr.*	MSS	Hang	3	na	na	na	na	na	na
	2004 Sheep Cr.*	MSS	Hang	5	na	na	na	na	na	na
	2003 Martin Cr.*	03Martin	Hang	13	na	na	na	na	na	na
	2004 Hangman Cr.*@	04Hang	Hang	9	na	na	na	na	na	na
	2003 Indian Cr.*	03Indian	Hang	37		0.75	9.05	x	0.053	0.004
	2004 Indian Cr.*	04Indian	Hang	21		0.77	na	na	0.091	<b>&lt;0.001</b>
(cutthroat) 2003 Nehchen Cr.*	03Nehc		65		0.60	8.51	x	0.162	<b>&lt;0.001</b>	
(cutthroat) 2004 Nehchen Cr.*	04Nehc		37		0.60	8.51	x	0.092	0.005	
Lower Hangman Cr.	2004 Marshall Cr.@	04Marsh		50		0.71	6.57	x	0.060	0.003
	2004 California Cr.@	04Calif		49	9	0.76	8.79	x	0.056	0.002
Spokane R.	2002 Middle Spokane R.*@	02MidSpok	Spok	47		0.77	10.49		0.056	0.002
	2003 Middle Spokane R.*@	03MidSpok	Spok	92		0.74	9.55		0.072	<b>&lt;0.001</b>
	2003 Upper Spokane R.*@	03UpSpok	Spok	67		0.67	7.62		0.071	<b>&lt;0.001</b>
	2004 Coulee Cr.*@	04Coulee		50		0.64	6.36		0.011	0.323
	2004 Deep Cr.*@	04Deep	Deep	50		0.73	8.17	x	0.035	0.031
	2004 Upper Deep Cr.*(@pond escape)	04UpDeep	Deep	50		0.70	6.71	x	0.018	0.209
Little Spokane R.	2003 Upper Little Spokane R.*@	03UpLSpok		39		0.75	9.07	x	0.043	0.023
	2003 Lower Little Spokane R.@	03LoLSpok		62		0.80	12.17		0.153	<b>&lt;0.001</b>
	2003 Dartford Cr. (above barrier)*	03DartAB	Dart	51		0.79	9.59	x	0.042	0.010
	2003 Dartford Cr. (below barrier)	03DartBB	Dart	29	3	0.78	10.39		0.001	0.477
	2003 Deadman Cr. @	03Dead	Dead	100	2	0.78	11.84		0.111	<b>&lt;0.001</b>
	2003 South Fork Deadman Cr. *	03SFDead	Dead	49	2	0.74	9.99	x	0.037	0.027
	2003 Little Deep Cr.	03LDeep	LDeep	50	1	0.75	11.00		0.030	0.040
	2003 North Fork Little Deep Cr.	03NFLDeep	LDeep	50	1	0.69	8.94		0.034	0.044
	2003 South Fork Little Deep Cr.	03SFLDeep	LDeep	60		0.71	9.46		0.062	<b>&lt;0.001</b>
	2001 Otter Cr.*@	01Otter		50	3	0.76	9.32	x	0.052	0.003
	2001 Deer Cr.*@	01DeerCr	Deer	100		0.77	11.86		0.033	0.003
	2002 Little Deer Cr.*	02LilDeer	Deer	50	1	0.73	10.34		0.037	0.028
	2001 Buck Cr.*@	01Buck		50		0.80	11.19		0.014	0.199
	2002 Lower Dragoon Cr. @	02LoDrag	Drag	100		0.79	11.56	x	0.047	<b>&lt;0.001</b>
2002 West Branch Dragoon Cr.	02WBDrag	Drag	50	2	0.77	9.53	x	-0.005	0.603	
2002 Upper Dragoon Cr.*	02UpDrag	Drag	50		0.77	10.16	x	0.022	0.112	
Hatch (Redband)	2001 Phalon Lake Hatch	01PhaH	PhaH	100	2	0.83	13.38	x	0.021	0.030
	2002 Phalon Lake Hatch	02PhaH	PhaH	100	1	0.79	11.57	x	0.081	<b>&lt;0.001</b>
Hatch (Uncertain)	2004 Trout Lodge Hatch	04TroutH		49		0.65	5.64	x	0.022	0.182
Hatch (Coastal)	2000 Spokane Hatch	00SpHat		96		0.70	6.35	x	-0.007	0.663
Hatch (Upper Columbia)	2002 Lyons Ferry Hatch steelhead	02LFH		100	2	0.74	11.70	x	0.065	<b>&lt;0.001</b>



1 Table 2. Sources and numbers of hatchery trout stocked into tributaries and lakes of the  
 2 Spokane drainage (from Brodie Cox, WDFW, unpublished data). Lakes under Buck  
 3 Creek were within the Buck Creek system.

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Tributary	# fish	Year planted	# Years planted	Source	
<b>Rainbow</b>					
Coulee Creek	7,000	1936	1	Spokane Hatchery	
Upper Little Spokane River	49,160	1940-1944	5	Spokane Hatchery	
Little Spokane River	922,240	1936 - 2002	36	Spokane Hatchery	
Little Spokane River	500	2002	1	Phalon Lake Hatchery	
Dragoon Creek	388,556	1934 - 1985	20	Spokane Hatchery	
Buck Creek	781,387	1941 - 1980	4	Spokane Hatchery	
	Diamond Lake	5,006,950	1933 - 1994	41	Spokane Hatchery
	Diamond Lake	69,768	1997	1	Lyons Ferry Hatchery
	Sasheen Lake	1,963,661	1939 - 1994	35	Spokane Hatchery
Deadman Creek	78,419	1934 - 1955	13	Spokane Hatchery	
Deer Creek	10,000	1936	1	Spokane Hatchery	
Otter Creek	10,000	1936	1	Spokane Hatchery	
Spokane River	1,348,134	1935 - 2002	34	Spokane Hatchery	
Spokane River	22,296	1996 - 1998	3	Phalon Lake Hatchery	
Spokane River	7,500	1995	1	Trout Lodge Hatchery	
Spokane River	16,886	1995	1	Lyons Ferry Hatchery	
Marshall Creek	628,911	1936-1983	25	Spokane Hatchery	
Deep Creek	24,448	1936-1944	4	Spokane Hatchery	
California Creek	10,000	1936	1	Spokane Hatchery	
Hangman Creek	3,573	1979 -1987	8	Spokane Hatchery	
<b>Cutthroat</b>					
Deep Creek	4,000	1939	1	Spokane Hatchery	
Dragoon Creek	19,000	1936 - 1937	2	Spokane Hatchery	
Little Spokane River and Buck Cr. area	5,116,802	1933 - 1992	36	Spokane Hatchery	
Marshall Creek	82,000	1957	1	Spokane Hatchery	

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1 Table 3. Information for multiplexes and loci including primer concentration, number of alleles in this study, size range (in  
 2 basepairs), observed heterozygosity ( $H_o$ ), repeat unit size (in basepairs),  $F_{ST}$ ,  $F_{IS}$ , and  $F_{IT}$  values (calculated without cutthroat trout, all  
 3 bold values significantly different from zero). References for primer sequences are under Source.  
 4

Multiplex	Anneal T	cycles	Locus	conc [uM]	#alleles	range	$H_o$	repeat	$F_{ST}$	$F_{IS}$	$F_{IT}$	Source
<i>Omy-B2</i>	55	26	<i>One-102</i>	0.05	34	182-305	0.789	4	<b>0.059</b>	<b>0.049</b>	<b>0.197</b>	Olsen <i>et al.</i> 2000
			<i>One-114</i>	0.05	26	177-280	0.828	4	<b>0.078</b>	0.024	<b>0.192</b>	Olsen <i>et al.</i> 2000
			<i>Ots-100</i>	0.04	26	168-298	0.826	2	<b>0.090</b>	-0.042	<b>0.096</b>	Nelson and Beacham 1999
<i>Omy-C2</i>	55	28	<i>One-101</i>	0.02	25	119-243	0.444	4	<b>0.115</b>	<b>0.098</b>	<b>0.395</b>	Olsen <i>et al.</i> 2000
			<i>One-108</i>	0.02	35	161-337	0.714	4	<b>0.088</b>	<b>0.181</b>	<b>0.298</b>	Olsen <i>et al.</i> 2000
			<i>Ots-103</i>	0.015	9	56-90	0.217	4	<b>0.099</b>	0.029	<b>0.100</b>	Small <i>et al.</i> 1998
<i>Omy-D2</i>	49	25	<i>Omy-77</i>	0.03	22	97-147	0.835	2	<b>0.068</b>	0.056	<b>0.246</b>	Morris <i>et al.</i> 1996
			<i>Ots-1</i>	0.03	22	158-266	0.727	2	<b>0.098</b>	0.014	<b>0.172</b>	Banks <i>et al.</i> 1999
			<i>Ots-3M</i>	0.02	11	132-156	0.573	2	<b>0.088</b>	<b>0.084</b>	<b>0.237</b>	Banks <i>et al.</i> 1999
<i>Omy-E2</i>	62	26	<i>Omm-1070</i>	0.025	42	164-374	0.797	4	<b>0.064</b>	<b>0.127</b>	<b>0.239</b>	Rexroad <i>et al.</i> 2001
			<i>Omm-1130</i>	0.05	57	185-399	0.800	4	<b>0.073</b>	<b>0.042</b>	<b>0.152</b>	Rexroad <i>et al.</i> 2001
			<i>Omy-1011</i>	0.045	21	134-245	0.798	4	<b>0.075</b>	<b>0.056</b>	<b>0.219</b>	Spies <i>et al.</i> 2005
<i>Omy-F2</i>	52	25	<i>Oki-10</i>	0.02	18	92-151	0.796	2	<b>0.095</b>	-0.060	<b>0.123</b>	Smith <i>et al.</i> 1998
			<i>Omy-1001</i>	0.03	30	167-242	0.810	2	<b>0.120</b>	-0.012	<b>0.147</b>	Spies <i>et al.</i> 2005

1 Table 4. Loci information for each collection:  $F_{IS}$  value at each locus, underlined values  
 2 were significant before Bonferroni correction and bold values significant after correction.  
 3

Gr. Spokane	One-102	One-114	Ots-100	One-101	One-108	Ots-103	Omy-77	Ots-1	Ots-3M	Omm-1070	Omm-1130	Omy-1011	Oki-10	Omy-1001
02MidSpok	0.096	0.078	0.032	0.182	0.026	0.076	0.018	0.08	-0.042	0.039	0.072	<u>0.125</u>	-0.003	-0.039
03UpSpok	0.056	0.049	0.012	<u>0.189</u>	0.089	-0.013	-0.053	<u>0.213</u>	-0.008	0.057	0.091	<b>0.316</b>	0.027	-0.015
03MidSpok	0.072	<u>0.113</u>	0.045	-0.126	0.059	-0.013	<u>0.093</u>	0.012	0.044	<b>0.176</b>	0.046	<u>0.149</u>	0.011	0.061
04Coulee	-0.020	-0.027	-0.031	-0.012	0.046	0.010	-0.072	-0.100	-0.114	<b>0.320</b>	0.020	0.101	-0.035	-0.031
04Deep	0.041	<u>0.213</u>	-0.001	0.150	0.106	-0.141	0.018	-0.013	0.004	<u>0.111</u>	-0.023	0.055	-0.017	0.006
04UpDeep	0.013	-0.034	-0.178	<u>0.351</u>	<b>0.436</b>	-0.044	-0.048	0.077	0.099	<u>0.196</u>	0.020	0.053	-0.075	-0.115
04Marshall	<u>0.206</u>	0.057	0.102	0.024	<b>0.268</b>	0.230	<b>0.342</b>	-0.010	<u>0.271</u>	0.013	-0.047	-0.095	-0.108	-0.080
04Calif	0.056	0.046	-0.014	-0.119	<u>0.205</u>	0.110	0.016	<u>0.159</u>	0.019	0.051	0.077	<u>0.143</u>	-0.041	<u>0.171</u>
MSS	<u>0.205</u>	<u>0.337</u>	0.106	-0.048	<u>0.175</u>	-0.034	0.211	<u>0.383</u>	0.222	-0.115	-0.119	0.153	-0.052	0.009
03Indian	-0.033	-0.058	-0.003	<u>0.383</u>	0.059	0.063	0.022	<b>0.560</b>	-0.093	<u>0.197</u>	<b>0.256</b>	0.029	0.055	-0.017
04Indian	<u>0.468</u>	0.042	-0.116	-0.250	0.029	<u>0.508</u>	-0.108	<u>0.416</u>	0.141	<b>0.337</b>	<u>0.207</u>	-0.116	-0.021	-0.054
04Hang	0.118	0.127	-0.032	0.289	0.097	0.000	-0.134	0.119	-0.347	-0.161	0.015	0.111	<u>0.453</u>	-0.134
03UpLSpok	-0.085	0.062	0.030	0.131	<u>0.177</u>	-0.118	0.082	0.052	0.011	0.061	0.093	0.111	0.044	-0.022
03LoLSpok	-0.027	<u>0.109</u>	0.047	0.068	<u>0.064</u>	0.148	0.012	<u>0.100</u>	-0.122	<b>0.726</b>	<b>0.544</b>	0.079	0.089	-0.010
03DartAB	0.014	0.095	-0.101	0.062	<u>0.153</u>	-0.130	0.017	-0.042	0.065	<b>0.231</b>	<u>0.129</u>	0.028	0.081	0.033
03DartBB	0.072	-0.022	-0.032	0.029	<u>0.289</u>	0.118	-0.106	-0.006	0.074	-0.042	0.103	-0.033	-0.092	-0.003
03Dead	<u>0.089</u>	<u>0.076</u>	<u>0.081</u>	0.068	<b>0.195</b>	-0.058	0.062	<b>0.194</b>	0.108	<b>0.256</b>	<b>0.196</b>	<u>0.072</u>	0.065	0.023
03SFDead	-0.049	-0.020	0.093	0.105	<b>0.361</b>	-0.055	-0.139	0.029	0.085	<u>0.154</u>	0.023	0.088	0.078	0.108
03LDeep	-0.130	0.012	0.028	-0.026	<u>0.193</u>	0.310	-0.025	0.102	0.036	0.004	0.087	<u>0.132</u>	<u>0.131</u>	0.028
03NFLDeep	0.079	-0.044	<u>0.218</u>	-0.028	<b>0.434</b>	-0.014	-0.045	0.099	-0.005	0.040	-0.010	0.027	0.006	-0.015
03SFLDeep	0.101	0.082	0.000	-0.152	<u>0.111</u>	-0.009	0.094	0.080	0.039	0.069	0.061	0.052	-0.014	<u>0.117</u>
01Otter	0.081	-0.004	0.047	-0.102	<b>0.273</b>	0.073	0.094	-0.020	<b>0.395</b>	0.054	0.087	-0.094	0.003	0.064
01DeerCr	0.042	0.019	0.038	0.084	<b>0.280</b>	0.198	0.037	-0.059	<u>0.198</u>	0.025	<b>0.110</b>	-0.063	0.032	-0.041
02LilDeer	-0.006	0.023	0.028	0.043	<b>0.371</b>	-0.088	0.064	0.053	0.031	<u>0.185</u>	<u>0.118</u>	-0.084	-0.042	-0.030
01Buck	0.045	-0.049	-0.003	0.042	<u>0.106</u>	-0.065	0.017	-0.061	0.043	<u>0.150</u>	-0.057	0.037	-0.016	0.024
02LoDrag	<u>0.105</u>	0.011	-0.033	0.019	<b>0.172</b>	-0.048	0.060	0.120	-0.036	<u>0.052</u>	<u>0.070</u>	0.062	-0.036	<u>0.082</u>
02WBDrag	-0.150	0.016	-0.029	0.133	<u>0.194</u>	0.045	-0.083	0.029	0.056	<u>0.139</u>	0.040	-0.008	-0.034	-0.153
02UpDrag	-0.077	-0.023	-0.003	-0.017	<u>0.165</u>	0.021	<u>0.191</u>	-0.026	-0.117	0.040	<u>0.123</u>	<u>0.128</u>	-0.114	-0.027
<b>Hatchery</b>														
01PhaHat	0.019	0.013	0.055	0.050	<u>0.084</u>	-0.004	-0.006	0.005	0.040	0.045	<u>0.054</u>	-0.026	-0.036	0.005
02PhaHat	<u>0.078</u>	0.048	0.020	-0.086	0.059	-0.126	0.037	<b>0.296</b>	0.012	<b>0.228</b>	<u>0.098</u>	0.030	-0.033	<u>0.070</u>
04TroutH	-0.013	-0.148	-0.155	0.124	0.115	0.000	0.080	-0.005	0.071	0.074	<u>0.267</u>	0.123	-0.129	-0.043
00SpHat	-0.034	0.043	-0.014	0.021	0.006	0.124	0.004	-0.006	0.020	0.004	-0.062	-0.018	-0.162	-0.106
02LFH	<u>0.061</u>	<u>0.058</u>	0.053	0.032	<u>0.113</u>	-0.085	0.052	<b>0.195</b>	0.007	<b>0.140</b>	<b>0.150</b>	-0.031	NA	0.010
<b>Cutthroat</b>														
03Nehc	-0.206	0.135	0.667	0.047	0.060	0.014	-0.242	0.108	0.064	<u>0.367</u>	<u>0.172</u>	<u>0.692</u>	-0.075	0.103
04Nehc	-0.289	-0.059	<u>0.365</u>	<u>0.774</u>	0.042	0.137	0.074	0.003	0.114	0.211	-0.016	0.407	0.021	-0.070

1 Table 5. Pairwise  $F_{ST}$  values with collections combined by tributary (see Table 1). Average pairwise  $F_{ST}$  values for wild collections  
 2 compared to other wild collections, and hatchery collections compared to only wild collections are under “Avg”. Highest average  
 3 value is in bold type. All pairwise values were significantly different from 0 ( $P < 0.0001$ ).  
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	Avg	Hang	Spok	04Coulee	Deep	04Marsh	04Calif	03UpLSpok	03LoLSpok	Dart	Dead	LDeep	01Otter	Deer	01Buck	Drag	PhaH	02LFH	00SpHat	TroutH
Hang	0.0912																			
Spok	0.0962	0.0566																		
04Coulee	<b>0.1647</b>	0.1493	0.1516																	
Deep	0.1102	0.1441	0.1275	0.1753																
04Marsh	0.1459	0.1815	0.1674	0.2387	0.1065															
04Calif	0.1000	0.0606	0.0739	0.1531	0.1480	0.1908														
03UpLSpok	0.1044	0.0899	0.1192	0.1972	0.1248	0.1702	0.1218													
03LoLSpok	0.0591	0.0629	0.0692	0.1439	0.0725	0.1075	0.0622	0.0664												
Dart	0.0859	0.1100	0.1145	0.1844	0.0817	0.1073	0.1110	0.0910	0.0475											
Dead	0.0628	0.0444	0.0580	0.1448	0.0973	0.1391	0.0614	0.0636	0.0279	0.0630										
LDeep	0.0816	0.0686	0.0851	0.1622	0.1164	0.1423	0.0930	0.0955	0.0499	0.0752	0.0360									
01Otter	0.0820	0.0871	0.0940	0.1705	0.0976	0.1459	0.0908	0.0878	0.0415	0.0664	0.0441	0.0676								
Deer	0.0695	0.0516	0.0654	0.1523	0.1156	0.1476	0.0560	0.0853	0.0224	0.0673	0.0194	0.0394	0.0512							
01Buck	0.0740	0.1013	0.0915	0.1673	0.0550	0.0867	0.0996	0.0821	0.0312	0.0364	0.0532	0.0676	0.0609	0.0668						
Drag	0.0623	0.0692	0.0725	0.1448	0.0802	0.1109	0.0774	0.0667	0.0228	0.0464	0.0271	0.0437	0.0423	0.0320	0.0362					
PhaH	0.0590	0.0566	0.0718	0.1371	0.0695	0.1048	0.0655	0.0608	0.0218	0.0523	0.0309	0.0556	0.0491	0.0397	0.0392	0.0301				
02LFH	0.0734	0.0546	0.0890	0.1460	0.1052	0.1431	0.0738	0.0693	0.0431	0.0742	0.0360	0.0613	0.0576	0.0405	0.0662	0.0417	0.0336			
00SpHat	0.1440	0.1868	0.1879	0.2336	0.1258	0.1616	0.1811	0.1357	0.1078	0.0982	0.1280	0.1436	0.1288	0.1436	0.0951	0.1019	0.1084	0.1465		
TroutH	0.1573	0.2011	0.1791	0.2668	0.1128	0.1388	0.2048	0.1796	0.1191	0.1163	0.1492	0.1628	0.1580	0.1663	0.0853	0.1247	0.1277	0.1623	0.1769	



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Table 6. Bayesian analysis of cutthroat-rainbow hybridization (clusters set at 2), and redband-coastal hatchery introgression (clusters set at 3). Values above 25% are in bold type.

	1	2		1	2	3	n
03UpSpok	<b>0.972</b>	0.028	03UpSpok	0.011	0.027	<b>0.963</b>	67
03MidSpok	<b>0.996</b>	0.004	03MidSpok	0.037	0.005	<b>0.959</b>	92
02MidSpok	<b>0.996</b>	0.004	02MidSpok	0.171	0.004	<b>0.825</b>	47
03UpLSpok	<b>0.996</b>	0.004	03UpLSpok	0.183	0.005	<b>0.813</b>	39
03LoLSpok	<b>0.996</b>	0.004	03LoLSpok	<b>0.387</b>	0.005	<b>0.608</b>	62
03DartAB	<b>0.998</b>	0.002	03DartAB	<b>0.450</b>	0.002	<b>0.548</b>	51
03DartBB	<b>0.998</b>	0.002	03DartBB	<b>0.508</b>	0.002	<b>0.489</b>	29
03Dead	<b>0.998</b>	0.002	03Dead	0.138	0.003	<b>0.859</b>	100
03SFDead	<b>0.997</b>	0.003	03SFDead	0.026	0.002	<b>0.971</b>	49
03LDeep	<b>0.998</b>	0.002	03LDeep	0.144	0.003	<b>0.853</b>	50
03NFLDeep	<b>0.998</b>	0.002	03NFLDeep	0.016	0.002	<b>0.982</b>	50
03SFLDeep	<b>0.998</b>	0.002	03SFLDeep	0.023	0.003	<b>0.975</b>	60
01Otter	<b>0.988</b>	0.012	01Otter	0.059	0.009	<b>0.932</b>	50
01DeerCr	<b>0.991</b>	0.009	01DeerCr	0.061	0.009	<b>0.930</b>	100
02LilDeer	<b>0.996</b>	0.004	02LilDeer	0.023	0.004	<b>0.973</b>	50
01Buck	<b>0.996</b>	0.004	01Buck	<b>0.750</b>	0.005	0.245	50
02LoDrag	<b>0.996</b>	0.004	02LoDrag	<b>0.267</b>	0.004	<b>0.729</b>	100
02WBDrag	<b>0.996</b>	0.004	02WBDrag	0.178	0.004	<b>0.818</b>	50
02UpDrag	<b>0.997</b>	0.003	02UpDrag	0.129	0.003	<b>0.868</b>	50
04Deep	<b>0.997</b>	0.003	04Deep	<b>0.981</b>	0.003	0.016	50
04UpDeep	<b>0.995</b>	0.005	04UpDeep	<b>0.983</b>	0.004	0.014	50
04Coulee	<b>0.998</b>	0.002	04Coulee	0.143	0.003	<b>0.855</b>	50
04Calif	<b>0.998</b>	0.002	04Calif	0.075	0.003	<b>0.923</b>	49
04Marsh	<b>0.998</b>	0.002	04Marsh	<b>0.970</b>	0.002	0.028	50
AllIndian	<b>0.989</b>	0.011	AllIndian	0.005	0.009	<b>0.985</b>	67
03Nehchen	0.025	<b>0.975</b>	03Nehchen	0.004	<b>0.971</b>	0.025	61
04Nehchen	0.003	<b>0.997</b>	04Nehchen	0.004	<b>0.991</b>	0.005	37
MSS	<b>0.998</b>	0.002	MSS	0.003	0.002	<b>0.995</b>	33
03Martin	<b>0.996</b>	0.004	03Martin	0.009	0.004	<b>0.987</b>	13
00SpHat	<b>0.997</b>	0.003	00SpHat	<b>0.993</b>	0.003	0.005	100
01Gold	<b>0.997</b>	0.003	01Gold	<b>0.987</b>	0.003	0.010	48
01Eell	<b>0.997</b>	0.003	01Eell	<b>0.993</b>	0.002	0.005	89
02STac	<b>0.997</b>	0.003	02STac	<b>0.988</b>	0.002	0.010	50
TroutH	<b>0.997</b>	0.003	TroutH	<b>0.991</b>	0.004	0.005	49
02LFH	<b>0.994</b>	0.006	02LFH	0.012	0.005	<b>0.982</b>	100
01PhaHat	<b>0.996</b>	0.004	01PhaHat	<b>0.295</b>	0.005	<b>0.700</b>	100
02PhaHat	<b>0.995</b>	0.005	02PhaHat	0.020	0.006	<b>0.973</b>	100
02SullCr	0.007	<b>0.993</b>	02SullCr	0.004	<b>0.990</b>	0.006	96
03SullLk	0.118	<b>0.882</b>	03SullLk	0.059	<b>0.871</b>	0.070	34
03GoldCr	0.019	<b>0.981</b>	03GoldCr	0.022	<b>0.973</b>	0.005	42

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1 Table 7. Bayesian analysis of shared ancestry in pairwise tests comparing a subset of  
 2 wild-origin rainbow trout collections from the Spokane River and Spokane Hatchery  
 3 rainbow trout.

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 5  
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 7

	<b>1</b>	<b>2</b>	n
00SpHat	0.005	<b>0.995</b>	100
04Marsh	<b>0.995</b>	0.005	50
00SpHat	0.014	<b>0.986</b>	100
01Buck	<b>0.988</b>	0.012	50
00SpHat	0.008	<b>0.992</b>	100
04Deep	<b>0.991</b>	0.009	50
04UpDeep	<b>0.995</b>	0.005	50
00SpHat	0.007	<b>0.993</b>	100
Drag	<b>0.988</b>	0.012	200
00SpHat	0.008	<b>0.992</b>	100
03DartAB	<b>0.994</b>	0.006	51
03DartBB	<b>0.987</b>	0.013	29
Spok	0.014	<b>0.986</b>	100
03LoLSpok	<b>0.991</b>	0.009	62
00SpHat	0.004	<b>0.996</b>	100
01Otter	<b>0.995</b>	0.005	50
00SpHat	0.003	<b>0.997</b>	100
04Coulee	<b>0.993</b>	0.007	50

8

1 Table 8. Maximum likelihood assignment test results with collections combined by tributary (see Table 1). Columns show the  
 2 number of fish from a single collection assigned to baseline collections. Positive, unambiguous assignments (100 times more likely  
 3 than assignment to another collection, no bracket around number) are followed by positive, ambiguous assignments (highest  
 4 assignment likelihood but less than 100 times more likely) in brackets. Fish assigned back to origin are along the diagonal. Total  
 5 number of fish in the baseline collection is in “total” row, number of correct assignments (unambiguous combined with ambiguous) is  
 6 in “correct” row, and percentage of correct assignments is in “% correct” row. Abbreviations follow Table 1.  
 7

	Hang	Spok	04Coulee	04Deep	04MarCr	04Calif	03UpLSpok	03LoLSpok	Dart	03Dead	03LDeep	01Otter	01DeerCr	01Buck	Drag	00SpHat	01PhaHat	02LFH	TroutH
Hang	105 (7)																		
Spok		188 (1)				(1)		(1)											
04Coulee		1	49	1					(1)										
04Deep		7 (2)	(1)	99							1 (1)								
04MarCr					50														
04Calif		(1)				45(2)													
03UpLSpok							39												(1)
03LoLSpok		(2)						38 (10)		3 (4)	(3)		(2)		(2)				(2)
Dart								2	76 (2)										
03Dead										104 (20)	(3)		(1)						
03LDeep	(1)							1 (1)		4 (5)	129 (21)								(2)
01Otter												49							(1)
01DeerCr		(1)						2 (5)		(1)			114 (31)						6 (3)
01Buck								(1)		(2)	(1)		(1)	47 (3)					(1)
Drag		(3)				(1)			(1)	(3)	(1)	(1)	(1)		169 (16)				
00SpHat																100			
01PhaHat								(1)		(3)							90 (7)		
02LFH																		99 (1)	
TroutH																			49
total	113	206	50	100	50	49	39	62	80	149	160	49	150	50	200	100	100	100	49
correct	112	189	49	99	50	47	39	48	78	124	150	50	145	50	185	100	97	100	49
% correct	99.12	91.75	98.00	99.00	100.00	95.92	100.00	77.42	97.50	83.22	93.75	102.04	96.67	100.00	92.50	100.00	97.00	100.00	100.00

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1 Figure 1. Map of Spokane River drainage showing tributaries where collections  
2 originated. Waterfalls are indicated by river kilometers in boxes. Man-made barriers and  
3 dry creek bed locations are designated approximately with triangles (Dartford Cr. barrier  
4 not shown). Filled circle marks the location of Spokane Hatchery. Map was constructed  
5 by Jim Shaklee (WDFW). City of Spokane (118°40'W; 47°50'N) is shaded with grey  
6 color.

7

8 Figure 2. Consensus neighbor joining tree of Cavalli-Sforza and Edwards genetic chord  
9 distances (1967) among rainbow and cutthroat collections from Spokane and Pend  
10 Oreille drainages and hatcheries. Numbers at the nodes indicate the percentage (greater  
11 than 60%) of 10,000 trees in which collections beyond the nodes grouped together.

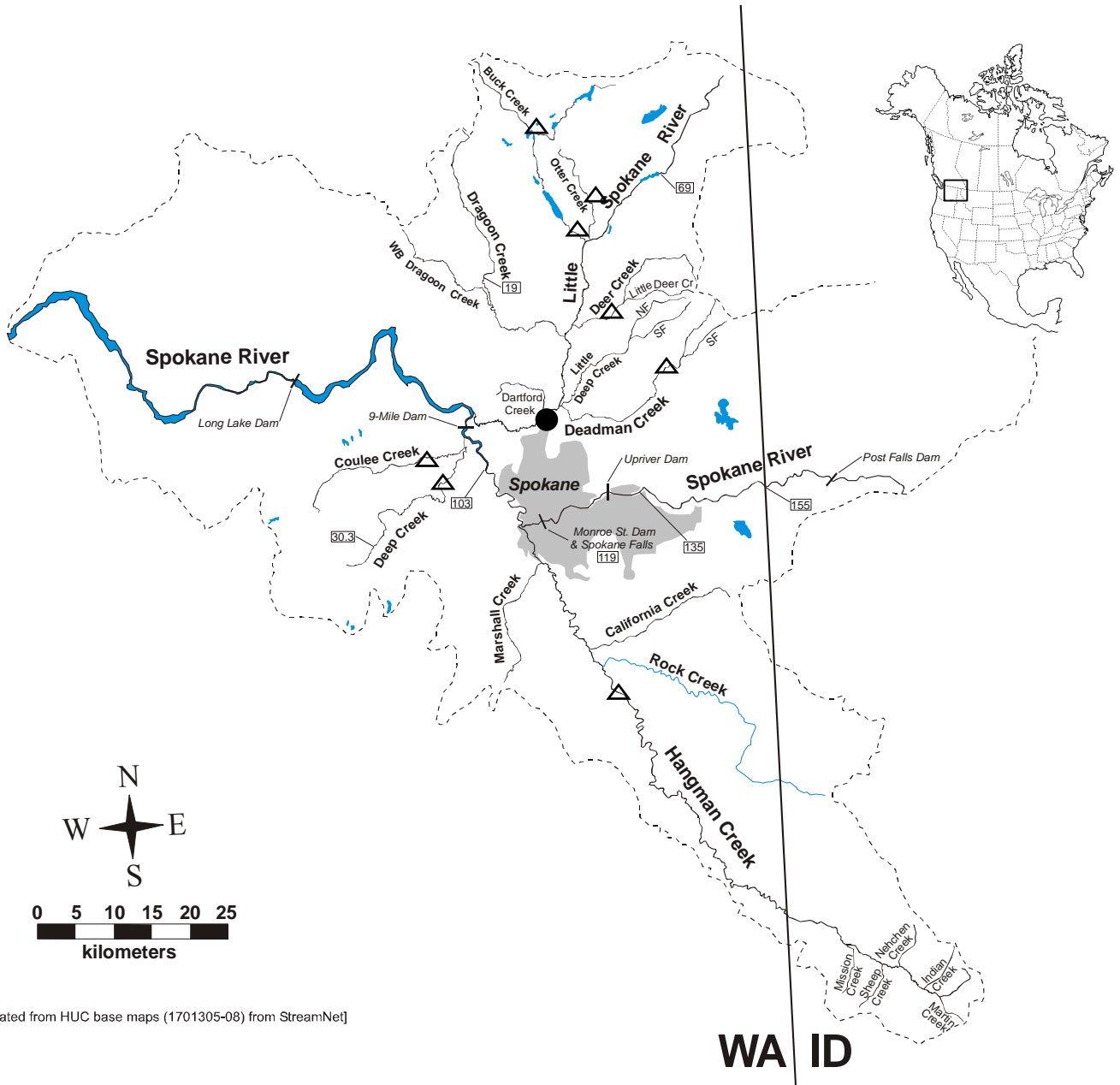
12

13 Figure 3. Multidimensional scaling analysis plot of Cavalli-Sforza and Edwards  
14 distances (1967) among hatchery collections and combined rainbow trout collections  
15 from the Spokane drainage.

16

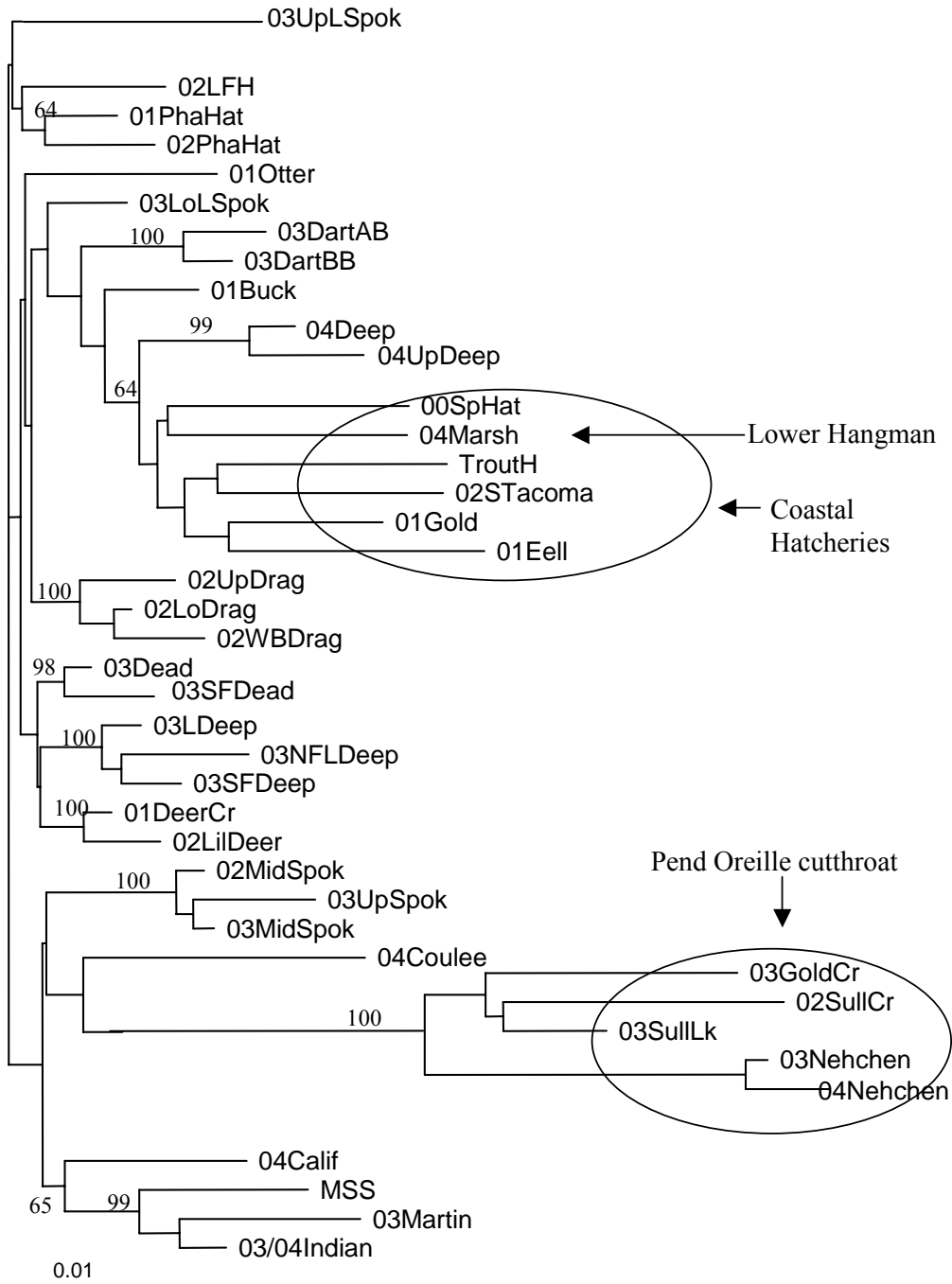
1 Figure 1.

2  
3



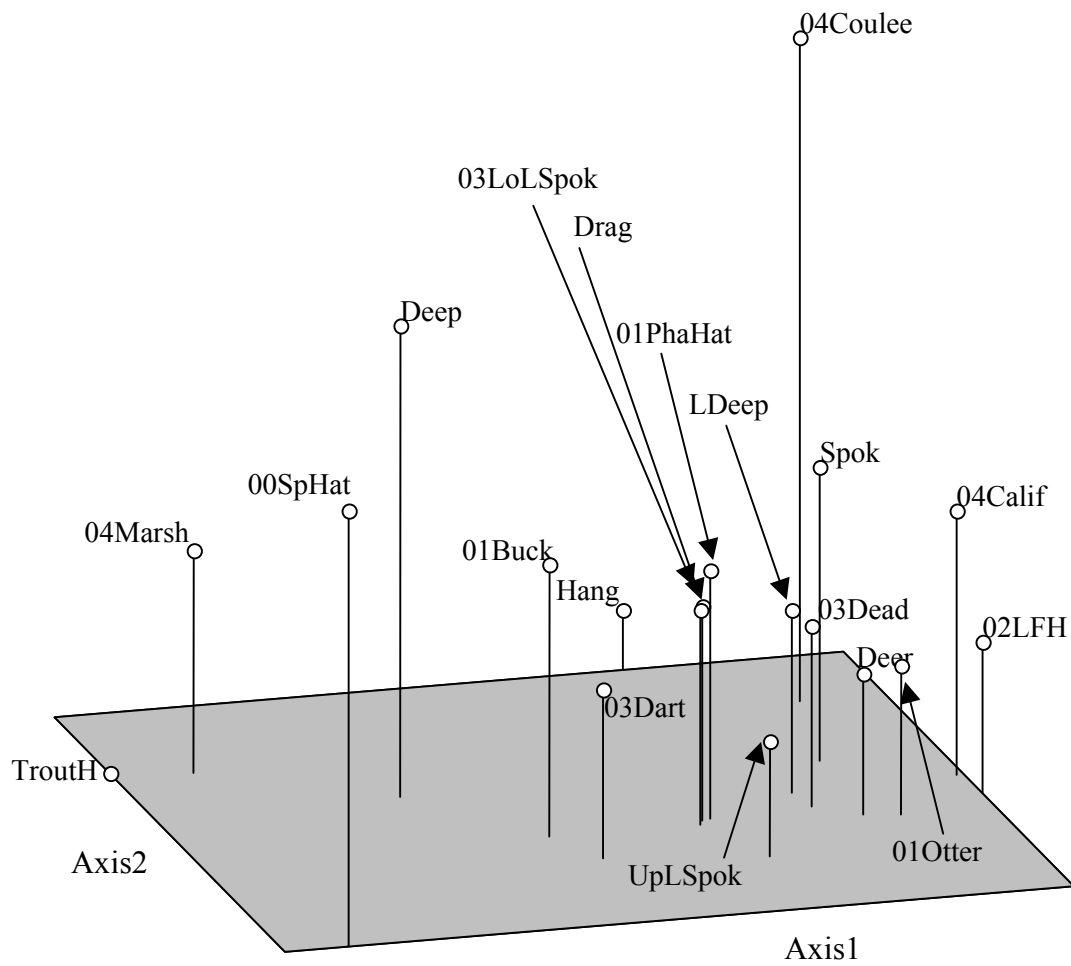
[map created from HUC base maps (1701305-08) from StreamNet]

1 Figure 2.  
2



3

1 Figure 3.  
2



3